

Microbial life under ice: metagenome diversity and *in situ* activity of Verrucomicrobia in seasonally ice-covered lakes

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Running title: Sub-ice Verrucomicrobia genomes in Quebec lakes

1 **Summary**

2 Northern lakes are ice-covered for a large part of the year, yet our understanding
3 of microbial diversity and activity during winter lags behind that of the ice-free period. In
4 this study, we investigated under-ice diversity and metabolism of Verrucomicrobia in
5 seasonally ice-covered lakes in temperate and boreal regions of Quebec, Canada using
6 16S rRNA sequencing, metagenomics and metatranscriptomics. Verrucomicrobia,
7 particularly the V1, V3 and V4 subdivisions, were abundant during ice-covered periods.
8 A diversity of Verrucomicrobia genomes were reconstructed from Quebec lake
9 metagenomes. Several genomes were associated with the ice-covered period and were
10 represented in winter metatranscriptomes, supporting the notion that Verrucomicrobia are
11 metabolically active under ice. Verrucomicrobia transcriptome analysis revealed a range
12 of metabolisms potentially occurring under ice, including carbohydrate degradation,
13 glycolate utilization, scavenging of chlorophyll degradation products, and urea use.
14 Genes for aerobic sulfur and hydrogen oxidation were expressed, suggesting
15 chemolithotrophy may be an adaptation to conditions where labile carbon may be limited.
16 The expression of genes for flagella biosynthesis and chemotaxis was detected,
17 suggesting Verrucomicrobia may be actively sensing and responding to winter nutrient
18 pulses, such as phytoplankton blooms. These results increase our understanding on the
19 diversity and metabolic processes occurring under ice in northern lakes ecosystems.

20 **Originality and significance**

21 Reduced ice cover on lakes is occurring worldwide, but there is only limited
22 information on the biogeochemistry and microbiology under ice. This gap in knowledge
23 limits our ability to understand and predict how changes in winter conditions will impact
24 the ecology of lake ecosystems. In this study, we conducted the first meta-omics
25 assessment of bacterial metabolism and gene expression patterns in seasonally ice-
26 covered lakes. Previously uncharacterized lineages within the Verrucomicrobia were
27 strongly associated with ice-covered conditions. Analysis of Verrucomicrobia genomes
28 and gene expression patterns revealed a diversity of functional trait including the use of
29 organic and inorganic energy sources and potential metabolic interactions with winter
30 phytoplankton and zooplankton. The identification of winter-associated lineages and
31 variable metabolic processes indicated that under-ice microbial communities may
32 contribute uniquely to the ecology and nutrient cycling of seasonally ice-covered lakes.
33 This study demonstrates the importance of studying the ice-covered period in the face of
34 climate change and should spur future year-round investigations on microbial community
35 structure and function in ice-covered freshwater ecosystems.

36

37

38 **Introduction**

39

40 Many northern lakes are seasonally ice-covered for over 40% of the year
41 (Weyhenmeyer *et al.*, 2011). During that time, these lakes are typically characterised by
42 low light levels and primary productivity, leading to the traditional idea that they are
43 “dormant” in winter (Bertilsson *et al.*, 2013). However, recent studies describing
44 phytoplankton and bacterial blooms as well as specialized microbial metabolism under
45 ice have challenged this concept (Kankaala *et al.*, 2006; Twiss *et al.*, 2012; Bertilsson *et*
46 *al.*, 2013; Bižić-Ionescu *et al.*, 2014; Beall *et al.*, 2016; Powers and Hampton, 2016). For
47 example, light penetrating through ice and snow has been shown to fuel substantial
48 primary production and blooms of low light and cold-adapted phytoplankton do occur
49 under the ice (Twiss *et al.*, 2012; Üveges *et al.*, 2012). These phytoplankton blooms can
50 fuel growth of other organisms, including heterotrophic bacteria (Bižić-Ionescu *et al.*,
51 2014). Moreover, unique niches can develop and persist in ice-covered lakes. Where ice-
52 cover results in lower availability of labile organic substrates from phytoplankton or
53 terrestrial input, volatile compounds such as methane and fermentation end-products
54 produced from anoxic bottom waters or sediments can fuel microbial metabolism (Sundh
55 *et al.*, 2005; Kankaala *et al.*, 2006). Chemolithoautotrophs may also be significant under
56 ice since their sources of energy and carbon acquisition may not be as restricted by ice
57 cover compared to phototrophs and heterotrophs (Auguet *et al.*, 2011). Despite these
58 observations, our understanding of the structure and function of microbial communities
59 under ice lags far behind that of the ice-free period (Bertilsson *et al.*, 2013; Powers and

60 Hampton, 2016; Hampton *et al.*, 2017). To better understand the contribution of winter
61 communities to lake metabolism and nutrient cycling, year-round investigation of the
62 metabolic traits and activities of freshwater microorganisms is warranted.

63 Verrucomicrobia are ubiquitous in lakes, yet knowledge of their metabolism and
64 ecology in freshwaters is limited compared to other bacterial groups (Newton *et al.*,
65 2011). The phylum is comprised of five orders, also referred to as subdivisions (V):
66 Verrucomicrobiae (V1), Spartobacteria (V2), Pedosphaera (V3), Opiritae (V4), and
67 Methyacidiphilum (V6) (Hedlund *et al.*, 1996; Sangwan *et al.*, 2004; Choo *et al.*, 2007;
68 Hou *et al.*, 2008; Kant *et al.*, 2011). 16S rRNA gene surveys have identified all
69 Verrucomicrobia subdivisions in lakes, however, only a few pelagic freshwater isolates
70 exist, all belonging to V1 (Hedlund *et al.*, 1996). In general, while Verrucomicrobia are
71 metabolically diverse, in aquatic ecosystems they are often associated with the
72 degradation of carbohydrates (Martinez-Garcia *et al.*, 2012; Herlemann *et al.*, 2013). For
73 example, analyses of single cell-amplified genomes (SAGs) and metagenome-assembled
74 genomes (MAGs) from coastal environments identified preferences for laminarin, xylan
75 (Martinez-Garcia *et al.*, 2012), cellulose and chitin polymers (Herlemann *et al.*, 2013).
76 Recently, comparative analysis of Verrucomicrobia MAGs from eutrophic Lake Mendota
77 and dystrophic Trout Bog in Wisconsin, USA showed differences in the number and type
78 of glycoside hydrolases (GHs) between the two systems, reflecting adaptations to local
79 environments and carbon substrate availability (He *et al.*, 2017). Additionally, MAGs
80 from freshwater reservoirs in Spain also suggested a preference for polysaccharides, but

81 also identified genes for the light driven proton pump rhodopsin as well as genes for
82 nitrogen fixation (Cabello-Yeves, Ghai, *et al.*, 2017). Perhaps most relevant from a life
83 under ice perspective, a recent study identified abundant populations of Verrucomicrobia
84 under the ice of Lake Baikal in Siberia, and these lineages also contained multiple
85 polysaccharide degradation pathways in their genomes (Cabello-Yeves, Zemskaya, *et al.*,
86 2017). Given their widespread distribution across freshwater ecosystems and their
87 prominence under ice, Verrucomicrobia may serve as a suitable model for investigating
88 metabolic adaptations and lifestyle strategies associated with seasonally ice-covered
89 lakes.

90 Here we present a study on Verrucomicrobia in seasonally ice-covered lakes in
91 Quebec, Canada with a general aim of advancing our understanding of the genomic
92 diversity and metabolic traits of bacterial communities residing under the ice of northern
93 lakes. Using a combination of 16S rRNA gene sequencing, metagenomics, and
94 metatranscriptomics we identified a wide diversity of Verrucomicrobia, including
95 populations and MAGs strongly associated with ice-covered periods. The MAG-
96 associated transcriptomes revealed a range of expressed metabolic genes under ice,
97 including those for the use of phytoplankton and plant-derived organic compounds,
98 chemolithotrophy, as well as motility and chemotaxis. Overall, our study supports the
99 increasing recognition that the winter period represents a dynamic and metabolically
100 important period for lakes despite ice cover.

101 **Results**

102 **Verrucomicrobia diversity and distribution in Quebec Lakes**

103 To investigate Verrucomicrobia abundance and diversity during ice-covered and
104 ice-free periods of the year, we analyzed a 3-year time-series of bacterial 16S rRNA gene
105 diversity from three seasonally ice-covered lakes located in the temperate (Lake Croche
106 and Lake Montjoie) and boreal (Lake Simoncouche) regions of Quebec (**Fig. 1A, Table**
107 **S1**). In total, we generated 16S rRNA data from 143 samples collected from the
108 epilimnion and metalimnion, and corresponding to 6 winter time-points (January and
109 February 2013, 2014, and 2015) and 8 summer time-points (June, July, August 2013 and
110 2014). The study lakes differed in several environmental characteristics, but were mainly
111 differentiated by nutrient concentrations in a principal component analysis (**Fig. 1B**). The
112 three lakes are distributed along a nutrient gradient, with Croche generally having lower
113 total phosphorus (TP) than Montjoie and Simoncouche (**Fig. S1, Table S2**).

114 Verrucomicrobia were common across all lakes, but generally exhibited higher
115 relative abundance of 16S rRNA gene sequences during ice covered period, particularly
116 in Croche and Simoncouche (**Fig. 1C**). The average relative abundance of
117 Verrucomicrobia was about 4-fold higher during ice cover (13%) compared to the ice-
118 free period (4%) in Croche and about 2-fold higher in Simoncouche (10% compared to
119 5%). Although more abundant under the ice on average, Verrucomicrobia were highly
120 dynamic in time with maximum observed values of 35.6 % and 27.1% in Croche and
121 Simoncouche during winter time points. Verrucomicrobia in Croche were dominated by

122 V1 and V3, while V4 was the major contributor in Simoncouche. In contrast to Croche
123 and Simoncouche, Verrucomicrobia exhibited a similar mean relative abundance between
124 ice-covered (4.8 %) and ice-free (6.2 %) periods in Montjoie. Maximum values in
125 Montjoie were observed during summer time points and were comprised mostly of V6.
126 Although relative abundance differed between lakes, these results demonstrate
127 Verrucomicrobia are often associated with ice-covered conditions, making them an
128 intriguing group with which to investigate genomic and metabolic adaptations to life
129 under ice.

130 **Verrucomicrobia MAG diversity and gene transcription**

131 Verrucomicrobia genome diversity was investigated in a metagenome co-
132 assembly of 24 samples collected over the 3-year times-series from the three lakes.
133 Following metagenomic binning, we identified 54 Verrucomicrobia MAGs representing
134 all 5 subdivisions (V1 to V4 and V6) previously observed in the 16S rRNA gene survey
135 (**Fig. 2A**). To investigate the distribution of MAGs across lakes and seasons, we
136 performed a Canonical Correspondence Analysis (CCA) of MAG coverage across
137 samples, constrained using 14 environmental variables (see methods for definitions of
138 abbreviations). In the resulting CCA, samples were differentiated along two primary axes
139 (**Fig. 2B**). Croche samples were separated from Montjoie and Simoncouche samples
140 along CCA axis 1, which correlated with nutrient concentrations (*i.e.* TP and TN). Ice-
141 covered samples were generally separated from ice-free samples along CCA axis 2,
142 which differentiated samples from the ice-covered and ice-free period, with the exception

143 of a single summer Montjoie sample, which clustered with winter samples. A clear
144 separation of MAGs along both axes of the ordination was observed, revealing lake and
145 seasonal preferences (**Fig. 2C**). Interestingly, MAGs within a subdivision, and even those
146 closely related within a subdivision, did not cluster together in the CCA, suggesting niche
147 diversity among members of the same subdivision; taxonomy clearly does not reflect
148 niche preference.

149 Fifteen MAGs of high completeness and low contamination were selected for
150 further analyses (**Table 1**). These MAGs represented four subdivisions (V1 to V4). The
151 MAGs were distributed across the CCA plot, and are therefore representative of the
152 observed phylogenetic diversity, biogeography, and seasonal associations of
153 Verrucomicrobia MAGs. A concatenated protein phylogeny that included 43 additional
154 MAGs from Wisconsin lakes (He *et al.*, 2017), Lake Baikal (Cabello-Yeves, Zenskaya,
155 *et al.*, 2017) and the Tous and Amadorio reservoirs in Spain (Cabello-Yeves, Ghai, *et al.*,
156 2017) was generated in order to place the Quebec MAGs in the context of known
157 Verrucomicrobia genomic diversity (**Fig. 3A**). The phylogeny presented in Figure 3A
158 was inferred from a concatenated alignment of four of the five proteins used in He *et al.*,
159 2017, but exhibited a similar topology to a phylogeny generated based on hundreds of
160 proteins (**Fig. S2**). MAGs were too distantly related to calculate average nucleotide
161 identity (ANI), but average amino acid identity between Quebec MAGs and those from
162 other locations ranged between 91-49% (**Table S3**). For the most part, Quebec MAGs
163 exhibited <70% AAI with other MAGs and therefore represent new lineages for which

164 genome sequence data is now available. Although common across Quebec lakes,
165 fragment recruitment of metagenomes originating from Wisconsin lakes, Lake Baikal,
166 and the Spanish reservoirs demonstrated that the Verrucomicrobia populations
167 represented by the MAGs are relatively rare in these other freshwater systems (**Table**
168 **S4**).

169 As revealed by their wide distribution in the CCA ordination, the MAGs exhibited
170 complex distribution patterns across Quebec lakes (**Fig. 3B**). To investigate gene
171 expression patterns of the Verrucomicrobia MAGs during ice-covered and ice-free
172 periods in Quebec, a temporally overlapping metatranscriptomic times-series was
173 mapped to the MAGs, providing a view of their transcriptional activity across lakes and
174 seasons. A significant number of transcripts were observed for all MAGs, and for the
175 most part reflected MAG distributions across lakes and seasons (**Fig. 3B**). In the
176 following section, we analyzed the distribution, metabolic gene content and gene
177 expression patterns of Verrucomicrobia MAGs with the objective of providing insights
178 into metabolic diversity and activity associated with ice-covered conditions in northern
179 lakes.

180 **Ice cover associated MAGs within V1**

181 A striking association with winter conditions was observed within certain V1
182 MAGs. Particularly, three MAGs (V1-159, V1-690, and V1-1361) were associated with
183 ice-covered conditions in Croche. These MAGs formed a clade with other freshwater
184 MAGs but were quite unique, sharing between 51-72 % AAI (**Fig 3A, Table S3**). V1-690

185 and V1-159 were exclusive to winter periods both at the level of the genome and the
186 transcriptome (**Fig. 3B**), while V1-1361 was also identified in the summer. A total of
187 3,983 orthologs were identified among these three MAGs and 868 were common to all
188 (**Fig. 4A**). Similar to other Verrucomicrobia, the winter-associated V1-159 and V1-690
189 MAGs contained numerous glycoside hydrolase (GH) genes (**Table S5**). Expression of
190 33 of 34 GH genes was detected in V1-159. The GHs of highest expression were GH29
191 (α -L-fucosidase) and GH16 (substrate specificity undetermined). In contrast,
192 expression was only detected for 12 of 54 GH genes in V1-690, suggesting other modes
193 of carbon and energy metabolism may be important for this population. Indeed, four
194 proteins annotated as monooxygenases were specific to V1-690 and were relatively
195 highly expressed. Two were annotated as limonene 1,2-monooxygenase and a third as 1
196 alkane 1-monooxygenase/p-cymene monooxygenase (**Fig. 4B**). Limonene and cymene
197 are plant-derived aromatic compounds, suggesting that these proteins are involved in
198 accessing terrestrial organic matter. In addition, we identified a genomic region encoding
199 the Sox system (*soxXYZABC*) together with two c-type cytochromes (**Fig. 4C**). The Sox
200 system is associated with the use of the reduced sulfur compound thiosulfate as an
201 electron donor in energy metabolism (Ghosh and Dam, 2009). All subunits were
202 expressed during winter except for *soxZ*, suggesting the potential for lithotrophic sulfur
203 oxidation during ice-covered periods.

204 **Ice cover associated MAGs within V4**

205 Specific winter association was observed for V4 MAGs. Two MAGs (Opi-128
206 and Opi-474) were phylogenetically related to MAGs from Trout Bog and the Tous
207 reservoir. Opi-128 exhibited 81 % AAI with TH02519 and TH01800 and Opi-474
208 exhibited 82 % AAI with TH4593, while Opi-128 exhibited 91 % AAI with Tous-
209 C10FEB (**Fig. 3A, Table S3**). Opi-474 and Opi-242 exhibited winter association in
210 Croche and Simoncouche, respectively (**Fig. 3B**). Opi-128 was more commonly
211 associated with summer. A total of 3,776 orthologs were identified among the three
212 genomes and only 11% were shared among all MAGs (**Fig. 4A**), indicating substantial
213 genomic diversity among V4 MAGs. Winter associated Opi-474 and Opi-242 contained
214 70 and 19 GH genes, respectively (**Table S4**). Compared to V1 winter MAGs, the
215 abundance of GH transcripts was elevated in the Opi-474 transcriptome. Among the
216 expressed GHs were GH78 genes involved in the use of rhamnase-containing
217 polysaccharides.

218 A common feature of Opi-474 and Opi-242 was the expression of a predicted
219 fatty acid α -oxidation II pathway (**Fig. 4D**). Fatty acid α -oxidation is implicated in the
220 metabolism of phytol, a long chain alcohol constituent of chlorophyll (Jansen and
221 Wanders, 2006). Phytol is first converted to phytanoyl-CoA, which then enters the α -
222 oxidation pathway where a methyl group at the C-3 position is removed before passage to
223 the more common β -oxidation pathway (Jansen and Wanders, 2006). Both genomes
224 encoded proteins of the phytanoyl-coA dioxygenase family, which hydroxylates the

225 methyl-branched fatty acid preparing it for downstream cleavage and passage to the β -
226 oxidation pathway. A putative pathway containing alcohol and aldehyde dehydrogenases
227 necessary for transforming phytol to phytanoyl-CoA was also identified. The presence of
228 a putative α -oxidation pathway and upstream steps for introducing phytol may allow
229 these organisms to scavenge carbon and energy from chlorophyll degradation products.

230 Evidence for chemolithotrophic energy conservation was observed in V4. A
231 complete gene cluster associated with aerobic hydrogen oxidation was specifically
232 identified in Opi-474, including genes encoding a Group 1d oxygen-tolerant hydrogenase
233 (Greening *et al.*, 2016) and associated maturation and nickel incorporation proteins (**Fig.**
234 **4C**). Expression of a number of these genes was detected under the ice in Croche,
235 suggesting the use of hydrogen as an electron donor in energy metabolism.

236 **Complex seasonal dynamics in V2 and V3 MAGs**

237 Compared to V1/V4 MAGs, V2 and V3 MAGs exhibited more complex
238 distribution patterns across lakes and seasons. In V3, Pedo-303 and Pedo-1123 branched
239 deeply within V3, and were present throughout the year in Montjoie, while Pedo-510
240 formed a clade with MAG recovered from Trout Bog and was present throughout the
241 year in Montjoie and Simoncouche (**Fig. 3**). V3 Quebec MAGs exhibited less than 66 %
242 AAI with MAGs from Trout Bog and the Tous reservoir.

243 The most complex seasonal patterns were observed for the four V2 MAGs. Two
244 MAGs (Xiphi-315 and Xiphi-554) were members of the *Xiphinematobacter*. Xiphi-315
245 was common but variable in Croche summer samples, while Xiphi-554 was restricted to

246 the June epilimnion in Simoncouche. Quebec *Xiphinematobacter* MAGs (Xiphi-315 and
247 Xiphi-554) were within the same clade as three MAGs from the Tous and Amadorio
248 reservoirs, but shared at most 63% AAI with them, reflecting high genome variation
249 between MAGs from Quebec lakes and other freshwater environments. Interestingly,
250 Xiphi-554 was the most widely distributed in other freshwater systems based on
251 metagenome fragment recruitment patterns (**Table S4**)

252 The two other V2 MAGs (Chth-244 and Chth-196) were members of the
253 *Chthoniobacter*. Chth-196 was the most broadly distributed MAG of all in Quebec lakes,
254 being common in Montjoie and Simoncouche but exhibiting high gene expression levels
255 in all. However, Chth196 was the only MAG that did not recruit any other freshwater
256 metagenome reads (**Table S5**), indicating very unique populations in Quebec lakes. Chth-
257 244 was common throughout the year in Montjoie although transcripts were only
258 detected during ice cover.

259 With respect to physiological adaptations to life under ice, Pedo-303 (V3) and
260 Chth-196 (V2) are of interest because differences in the number of transcripts between
261 summer and winter was detected for both MAGs, providing insight into metabolic
262 responses to different seasons. Although Pedo-303 and Chth-196 were distantly related
263 from a phylogenetic perspective, they exhibited some similarities at the metabolic level.
264 For example, ammonia and urea appear to be important nitrogen sources for both groups.
265 Urea transport proteins and urease subunits were identified in Chth-196 and generally had
266 higher expression during winter than summer in Simoncouche. Urea transport and urease

267 genes were also present in Pedo-303 and more transcripts were detected during ice cover
268 than the ice-free period in Montjoie (**Fig. 5C**). Furthermore, Pedo-303 had genes
269 implicated in ammonia transport and assimilation, including three copies of the nitrogen
270 regulatory proteins P-III, two Amt family ammonium transporters, and two glutamine
271 synthetases, which had higher levels of expression in the winter (**Fig. 5D**). In Chth-196,
272 three Amt family transporters, one glutamine synthetase, and three nitrogen regulatory
273 proteins P-II-1 were found, and these genes were expressed more in Simoncouche during
274 ice cover, but more in Montjoie during the summer.

275 Pedo-303 and Chth-196 both encoded genes for the use of a range of organic
276 carbon compounds. In Pedo-303, rhamnose and xylose utilization genes were expressed
277 at both time points, but transcript abundance was higher in the winter. In Chth-196, genes
278 for rhamnose and ribose transport and degradation were identified and expressed during
279 winter. The use of glycolate, a photorespiration product of phytoplankton, was also
280 suggested by the presence of a glycolate utilization operon (Fe-S subunit, FAD-binding
281 subunit, oxidase) in Pedo-303. Transcripts for glycolate use were only detected during
282 winter in Montjoie.

283 Another notable observation within V2 and V3 MAGs was the relatively high
284 number of transcripts for genes involved in motility and chemotaxis. Genes for flagellar
285 biosynthesis were identified in numerous MAGs (Chth-196, Chth-244, Opi-474, Pedo-
286 510 and Xiphi-554), but evidence for active use of the flagellar machinery was found in
287 Chth-196 and Pedo-510 only. In Chth-196, transcripts encoding flagellar motility proteins

288 were generally more abundant in summer in Montjoie, but in winter in Simoncouche
289 (**Fig. 5A**). In addition, Chth-196 encoded genes involved in chemotaxis behaviour were
290 more abundant in winter than summer in Simoncouche (**Fig. 5B**).

291 **Discussion**

292 In this study, we conducted the first meta-omics assessment of bacterial
293 metabolism and gene expression patterns in seasonally ice-covered lakes. In combination
294 with other recent metagenomic studies (Cabello-Yeves, Ghai, *et al.*, 2017; Cabello-
295 Yeves, Zenskaya, *et al.*, 2017; He *et al.*, 2017), our findings contribute to the emerging
296 view that Verrucomicrobia is an important, but previously overlooked component of lake
297 microbial ecosystems. The association of Verrucomicrobia MAGs with ice-covered
298 conditions suggests certain populations exhibit a winter preference. However, an
299 alternative explanation for greater under-ice abundance is that Verrucomicrobia prefer as
300 habitat the deeper hypolimnetic regions of northern lakes, and thus they could appear
301 winter-associated following autumn mixing of the water column. Although we cannot
302 rule out this possibility, it seems unlikely given that sampling occurred several months
303 following the onset of fall overturn. Furthermore, we did not observe evidence for
304 hypolimnetic preferences via anaerobic metabolism in the Verrucomicrobia MAGs
305 despite the fact that Croche and Simoncouche have anoxic hypolimnia during summer.
306 However, Verrucomicrobia have been observed in the oxic hypolimnion of the Tous and
307 Amadorio reservoirs (Cabello-Yeves, Ghai, *et al.*, 2017). Moreover, studies of the sub-ice

308 microbial communities of deep Lake Baikal have shown Verrucomicrobia are among the
309 most dominant groups at the surface (Cabello-Yeves, Zemskaya, *et al.*, 2017) and that
310 their abundances are linked to diatom blooms (Bashenkhaeva *et al.*, 2015). Hence, it
311 appears that Verrucomicrobia populations represented by our MAGs are resident
312 members of the microbial community in ice-covered periods in northern lakes, and
313 knowledge of their genomic and metabolic traits can contribute to our understanding of
314 lake metabolism and nutrient cycling in winter.

315 Ice-covered conditions may favour microbes that have low resource requirements
316 (*i.e.* oligotrophs) and small genome size is a common adaptation to oligotrophic
317 conditions (Carini *et al.*, 2012; Neuenschwander *et al.*, 2017). Previous analysis of
318 Verrucomicrobia MAGs have shown a wide range in genome sizes (Cabello-Yeves, Ghai,
319 *et al.*, 2017). We speculated that winter-associated MAGs may be specialized to low
320 carbon/energy conditions and that this would manifest in overall smaller genomes
321 compared to those associated with the ice-free period. However, we did not observe
322 generalizable differences in estimated genome size across seasons within Quebec lakes.
323 Rather, we observed that overall estimated genome size was on average smaller in
324 Verrucomicrobia MAGs from more oligotrophic lakes (Quebec, Lake Baikal, and
325 freshwater reservoirs) compared to eutrophic lakes (Mendota and Trout Bog). Wisconsin
326 MAGs had the largest average estimated genome size in Subdivision 1, 2 and 3 compared
327 to Quebec, Baikal, Tous and Amadorio MAGs. The largest average estimated genome
328 size in V4 was in Tous and Amadorio MAGs, while the smallest was among Quebec

329 MAGs. Based on our results, small genome size does not seem to be distinguishing
330 characteristic of winter association, at least for Verrucomicrobia, but instead reflects
331 adaptations to oligotrophic freshwater environments in general.

332 **Verrucomicrobia-phytoplankton coupling**

333
334 Phytoplankton can persist and form transient blooms during ice cover (Twiss *et*
335 *al.*, 2012) and a number of studies have implicated Verrucomicrobia in the degradation of
336 phytoplankton-derived carbohydrates (Paver and Kent, 2010; Parveen *et al.*, 2013),
337 including in ice-covered lakes (Bižić-Ionescu *et al.*, 2014; Beall *et al.*, 2016). For
338 example, Verrucomicrobia were among the main bacteria associated with diatom-
339 dominated under ice blooms in Lake Baikal (Bashenkhaeva *et al.*, 2015). Furthermore,
340 Verrucomicrobia OTUs, including *XipA1/ XipB1* (V2) and *Opitutacea* (V4) were strongly
341 correlated with algal carbon in Finnish subarctic lakes (Roiha *et al.*, 2016). Finally,
342 Verrucomicrobia (*Verrucomicrobiae* (V1) and *Opitutacea* (V4)) abundance significantly
343 increased following the release of extracellular polymeric substances released by diatoms
344 in intertidal zones (Bohórquez *et al.*, 2017)

345 Numerous traits of the Quebec lake Verrucomicrobia suggest a capacity to couple
346 growth to phytoplankton, including winter phytoplankton blooms. Based on 16S rRNA
347 gene analysis, Verrucomicrobia subdivisions were commonly abundant in both the free-
348 living and particle-attached fractions of the community, indicating an ability to switch
349 between these lifestyles. During winter, cells may persist in a free-living state but also

350 associate with phytoplankton during bloom onset and progression. If so, then the MAGs
351 analysed here originated from the free-living persisters. Interestingly, GH genes were
352 identified in all MAGs, with no strong correlation between estimated genome size and
353 number of GH found. However, we found a wider diversity of GHs was expressed in
354 winter compared to summer, perhaps in readiness for persister cells to quickly respond to
355 carbohydrate that becomes available, both from phytoplankton or terrestrial sources.
356 Expression of genes for the use of phytoplankton-derived saccharides including fucose
357 and rhamnose were identified, suggesting that Verrucomicrobia interact with
358 phytoplankton during winter. Future studies comparing gene expression patterns between
359 particle-attached and free-living cells, would be informative in understanding the
360 metabolic shifts that occur during these transient bloom events.

361 Intriguingly, expression of a glycolate degradation operon was detected in ice
362 cover associated MAGs from Lake Montjoie. Glycolate is a photorespiration product
363 (Fogg, 1983) that is typically produced by phytoplankton growing under high light stress
364 (Parker and Armbrust, 2005; Davis *et al.*, 2013). Glycolate-utilizing bacteria have been
365 shown to be transcriptionally responsive to phytoplankton blooms (Lau *et al.*, 2007).
366 Glycolate production during winter may be unexpected owing to low light penetration
367 through snow and the absence of light stress (Maxwell *et al.*, 1994). However, low-light
368 adapted phytoplankton may have a lower tolerance to light overall. Rapid increases in
369 light intensity due to blowing/melting snow could lead to light stress, and a
370 corresponding pulse of glycolate into the environment. Another possibility is that

371 glycolate produced during the ice-free period is relatively long-lived in the water column.
372 Glycolate is present in the ocean year-round, and can account for as much as 33% of the
373 dissolved organic carbon pool (Carlson and Ducklow, 1996; Lebourlanger *et al.*, 1997). In
374 any case, it is believed that the main role of glycolate in heterotrophic metabolism is as an
375 energy source (Wright and Shah, 1977; Edenborn and Litchfield, 1987). Recently, a study
376 of carbon utilization in temperate lakes showed that a relatively larger portion of
377 phytoplankton-derived carbon was allocated to respiration (and hence energy
378 conservation), while terrestrial carbon was allocated to biosynthesis (Guillemette *et al.*,
379 2016). Therefore, glycolate may be an important contributor to the maintenance energy of
380 a persisting cell or could facilitate growth by serving as an energy source, fuelling the
381 subsequent assimilation of terrestrial carbon in winter.

382 Phytoplankton blooms and resource availability under ice are often patchy.
383 Motility and chemotaxis may be useful adaptations to efficiently exploit hotspots of
384 organic and inorganic nutrients (Stocker and Seymour, 2012). This idea is supported by a
385 number of Verrucomicrobia MAGs that expressed genes for flagella and chemotactic
386 abilities. Of particular interest were two closely related winter MAGs from Montjoie
387 (Chth-244) and Simoncouche (Chth-196). Flagella and chemotaxis gene expression was
388 evident in the Montjoie MAG. Despite the presence of flagella and chemotaxis genes in
389 the Simoncouche MAG, only the flagella genes were detectably expressed. The
390 difference in chemotactic activity may reflect the difference in lake physical conditions;
391 while Montjoie is unstratified in winter, Simoncouche is inversely stratified. Hence, the

392 ability to detect and respond to chemical gradients may be more advantageous in
393 stratified systems. Operating at the micro-scale rather than meters, this notion of motility
394 towards hotspots of nutrients is analogous to findings in phytoplankton motility. For
395 example, in a 20-year study of phytoplankton functional traits in an ice-covered lake in
396 Germany, authors found that during mild winters a mixed water column favoured non-
397 motile phytoplankton (Özkundakci *et al.*, 2016), as in Montjoie. Therefore, the
398 difference in water column mixing between lakes is potentially reflected in bacterial
399 functional traits as well.

400 **Scavenging of the phytol moiety of chlorophyll**

401 In addition to responding to algal growth, bacterial scavenging of dead and
402 degraded phytoplankton and other detritus may be an important metabolic strategy for
403 life under ice. Labile algal carbohydrates and proteins would likely be the first
404 biomolecules scavenged from the environment, but carbon-rich lipids may also sustain
405 growth during winter. In freshwater reservoirs, Verrucomicrobia metagenomes contained
406 several enzymes (aryl sulfatases, beta-galactosidases and sialidases) for degradation of
407 glycosphingolipids and are suggested to be involved in the degradation of plant or algae
408 derived sulfur-containing lipids (Cabello-Yeves, Ghai, *et al.*, 2017). In our study, winter
409 MAGs encoded for a predicted α -oxidation pathway. The α -oxidation pathway is
410 implicated in the degradation of phytol, the long-chained alcohol moiety of chlorophyll.
411 Early studies showed growth of bacteria on phytol as a sole source of carbon and energy
412 (Hoag *et al.*, 1969), however the ecological significance of this metabolism is unknown.

413 The concentration of free phytol in the water column may be low, but we can envision
414 microenvironments where phytol concentrations are elevated. One micro-environment
415 may be formed by the fecal pellets of zooplankton. Early studies have found that the fecal
416 pellets of zooplankton, specifically coccoliths feeding on an algal-rich diet, were rich in
417 phytanic acid, the end product of phytol oxidation that feeds into the α -oxidation pathway
418 (Moussa, 1988). The phytanic acid-rich particles would be available in the water column
419 for consumption by bacteria possessing the α -oxidation pathway

420 Additionally, it is possible that phytol may accumulate in particles that sink into
421 the anoxic hypolimnion of lakes. Since the α -oxidation pathway requires oxygen (through
422 the essential phytanoyl CoA hydroxylase), phytol may accumulate under anoxic
423 conditions, although anaerobic degradation of phytol in sediments has been reported
424 (Rontani *et al.*, 1999). Nevertheless, during breakdown of stratification in autumn, phytol
425 accumulated in the hypolimnion could serve as a carbon reserve supporting microbial
426 metabolism throughout the winter. A recent study in Lake Simoncouche has shown that
427 zooplankton are able to store fatty acids from phytoplankton to survive over the winter
428 (Grosbois *et al.*, 2017). Similarly, bacteria might be able to use zooplankton and
429 phytoplankton-derived lipids (e.g. phytanic acid) to maintain baseline metabolism under
430 ice. These observations point to the complex relationships among bacteria,
431 phytoplankton, and zooplankton occurring under the ice in freshwater lakes.

432 **Chemolithotrophy under ice**

433 Chemolithotrophic growth is traditionally thought to be restricted to sediments or
434 the chemocline of stratified lakes where there is a sufficient supply of reduced inorganic
435 compounds for metabolism. Here we detected the expression of sulfur and hydrogen gene
436 clusters, suggesting that chemolithotrophic energy metabolism may be employed by
437 bacteria under the ice. Similar sulfur oxidation genes were also recently reported in *Beta-*
438 *proteobacteria* MAGs from Lake Baikal (Cabello-Yeves, Zemskaia, *et al.*, 2017).
439 Heterotrophic organisms capable of supplementing their energy demand using inorganic
440 compounds would have an advantage over those that cannot if the availability of organic
441 carbon is limited in winter. A wide diversity of so-called “heterotrophic sulfur-oxidizing”
442 bacteria are found in marine systems (Teske *et al.*, 2000). In these cases, sulfur oxidation
443 allows bacteria to allocate organic carbon for biosynthesis instead of respiration, allowing
444 them to thrive in a wider range of habitats (Teske *et al.*, 2000; Podgorsek *et al.*, 2004).
445 Similarly, the winter water column availability of H₂ (previously restricted to the
446 anaerobic hypolimnion), combined with the availability of oxygen under ice, may allow
447 aerobic hydrogen-oxidizing to use this energy rich fermentation end-product for growth
448 or maintenance energy.

449 **Urea as an important N source in lake**

450 The observation of urea utilization genes in our MAGs further suggests the
451 potential for trophic interactions. Although urea is usually present in lakes at ambient

452 concentrations below 1 $\mu\text{M-N}$, it can contribute 50% or more of the total N used by
453 planktonic communities (Solomon et al., 2010). Interestingly, urea availability is
454 predicted to be regulated mainly by the decomposition of algae under anoxic conditions
455 (e.g. hypolimnion/sediments), followed by redistribution in the water column (Bogard *et*
456 *al.*, 2012) and thus potentially serving as a valuable nitrogen source in winter. All urease
457 subunits were found and expressed in the winter in Chth-196 and Pedo-303. A recognized
458 source of N for polar phytoplankton, the importance of urea for bacteria has received less
459 attention. In the arctic, urea utilization was detectable in the prokaryotic size fraction, but
460 not that corresponding to the phytoplankton fraction, showing the potential importance of
461 urea for cold-adapted metabolism, and adaptation to low energy environments (Alonso-
462 Sáez *et al.*, 2012). In Lake Baikal, urea utilization genes were found in *Cyanobacteria*,
463 *Acidobacteria*, *Nitrospira*, *Beta-proteobacteria* and *Thaumarchaeota* MAG (Cabello-
464 Yeves, Zemskaya, *et al.*, 2017). The expression of the urea utilization genes by
465 Verrucomicrobia that we observed suggests that urea can be beneficial for bacterial
466 species living under the ice in temperate and boreal lakes as well, given the similar
467 constraints of temperature, light and organic nutrient availability.

468

469 **Implications**

470 There is a growing acknowledgment among aquatic ecologists as to the need to
471 study the full annual cycle of lakes in order to understand lake dynamics, particularly as
472 lake temperature increases worldwide (O'Reilly *et al.*, 2015). This is exemplified by a

473 recent quantitative synthesis on under ice ecology in which winter-summer patterns of
474 nutrients, phytoplankton, and zooplankton were investigated (Hampton *et al.*, 2017)
475 Though full-year time-series of these variables are rare, studies that also include
476 microbial community structure and function variables are even more so although
477 impressive long-term microbial time-series of the ice-free period do exist (Bendall *et al.*,
478 2016; Linz *et al.*, 2017). In addition to a few other studies, our findings substantiate the
479 idea that understanding the under-ice microbiome is important for predicting the
480 dynamics of seasonally ice-covered lakes in the future (Bertilsson *et al.*, 2013). Although
481 this study was illuminating regarding under ice microbial metabolism, we focused solely
482 on winter-associated Verrucomicrobia. Community wide winter-summer patterns remain
483 to be elucidated and will likely vary tremendously between lakes. Hence, a much wider
484 range of studies must be executed before generalizable patterns can be reported.
485 Moreover, meta-omics studies such as this are, in essence, hypothesis-generating and
486 future work that includes targeted enrichment/cultivation and *in situ* rate measurement-
487 based approaches are required to validate and quantify microbial metabolic contributions
488 to nutrient cycling in lake environments during winter.

489 **Methods**

490 **Sampling**

491 Microbial samples were collected from three freshwater lakes: Lake Croche
492 (45°59'N, 74°01'W), Lake Montjoie (45°24'N, 72°14'W) and Lake Simoncouche
493 (48°14'N, 71°15'W) in conjunction with the Groupe de Recherche Interuniversitaire en
494 Limnologie et en Environnement Aquatique (GRIL) Monitoring Program. During 3 years
495 (2013-2015), epilimnion and metalimnion samples were collected biweekly during ice-
496 free periods and monthly during ice-covered periods of the year (see **Table S1** for a full
497 description of samples collected for microbial analyses). In winter, samples were
498 collected from just below the ice, as well as deeper in the water column for lakes
499 Simoncouche and Croche, while a single integrated sample of water collected from
500 multiple depths was collected from Montjoie. Lake water, collected in acid-washed
501 bottles, was pre-filtered through 53 µm mesh, followed by sequential filtration on to a 3
502 µm polycarbonate filter to collect particle-attached cells followed by a 0.22 µm Sterivex
503 filter to collect free-living cells. 1.8 ml of sucrose-based lysis buffer was added to
504 samples collected for DNA extraction, 1.8 ml of RNA later was added to samples
505 collected for RNA extraction, and filters were stored at -80° C until processing.

506 Water column profiles of environmental variables (**Table S2**) were measured
507 including temperature, dissolved oxygen, pH, specific conductivity and oxidation
508 potential, directly in the field using a multiparameter sonde (YSI, OH, USA). Water
509 samples were collected at the same depths as the microbial samples. Analyses of total

510 phosphorus (TP), total nitrogen (TN), total dissolved phosphorus (TDP), total dissolved
511 nitrogen (TDN), dissolved organic carbon (DOC), nitrate (NO_3^-), nitrite (NO_2^-),
512 ammonium (NH_4^+), ions (Cl^- , PO_4^{3-} , SO_4^{2-} , Ca^{2+} , Mg^{2+} , Na^+), and Chlorophyll *a* were
513 performed in the GRIL laboratory at Université de Montréal (Montreal, Canada).

514 **Nucleic acid extraction**

515 DNA was extracted from 0.22 μm Sterivex filters using a phenol/chloroform-
516 method modified from Zhou (1996). Sterivex filters were thawed on ice and the storage
517 buffer was removed. The storage buffer was concentrated into Amicon 30 kD filter (500
518 μl at a time) followed by centrifugation for 20 minutes at 10,000 g. 500 μl storage buffer
519 was repeatedly added until concentrated to a final volume of 100 μl . Buffer exchange was
520 conducted twice by washing storage buffer with 500 μl of TENP buffer (600 mg Tris,
521 740 mg EDTA, 580 mg NaCl, 2 g Polyvinylpyrrolidon and 100 ml milliQ, pH 8). We
522 then broke open the Sterivex filter and removed the filter. The filter was split into halves
523 and placed inside a 2 ml Eppendorf tube. To conduct the cell lysis and digestion, 0.37
524 grams of 0.7 mm pre-sterilized Zirconium beads, 60 μl of 20 % SDS, 100 μl concentrated
525 buffer exchanged filtrate, 500 μl TENP buffer and 500 μl phenol-chloroform-
526 isoamylalcohol (PCI) 25:24:1 were added to the 2 ml Eppendorf tube containing the
527 shredded filter. Then, the sample was vortexed for 10 minutes. Samples were incubated
528 for 10 min in a 60 °C water bath followed by incubation on ice for 1 min. The samples
529 were centrifuged for 6 min at 10,000 rpm and 4°C. Supernatant was transferred to a clean
530 1.5 ml Eppendorf tube, 500 μl phenol-chloroform-isoamylalcohol (PCI) 25:24:1 was

531 added and samples were vortexed briefly. Then, samples were centrifuged for 6 min at
532 10,000 rpm and 4 °C. The supernatant was transferred to a new 1.5 ml Eppendorf tube.
533 The PCR step was repeated until there was no longer any white precipitate at the interface
534 (usually 2 times). DNA was precipitated by adding 120 µl of 3 M sodium acetate
535 followed by 1 ml of 96% ethanol. DNA was precipitated at -20 °C for at least 1.5 hours,
536 followed by centrifugation for 60 minutes at 13,000 rpm and 4 °C. The supernatant was
537 decanted, and the pellet was washed with 850 µl of 80 % ethanol. We incubated samples
538 for 10 minutes on ice followed by short vortexing and then centrifuging samples for 15
539 minutes at 13,000 rpm and 4 °C. The supernatant was removed and the DNA was
540 resuspended in 50 µl TE or Tris-HCl (pH 7.5-8).

541 RNA was extracted from 0.22 µm Sterivex filters with a modified protocol (Shi *et*
542 *al.*, 2009; Stewart *et al.*, 2010) which employs both the mirVana miRNA isolation kit
543 (Invitrogen) and the RNeasy RNA cleanup kit (Qiagen). Samples were thawed and had
544 the RNAlater (Invitrogen) surrounding the Sterivex filter removed (approximately 1700
545 µl) and discarded. 1700 µl of mirVana lysis buffer was added to the Sterivex filter and
546 vortexed to lyse bacterial cells attached to the filter. Total RNA was then extracted from
547 the lysate according to the mirVana protocol. Purified sample (100 µl) was treated with 2
548 µl DNase (New England Biotech) incubated at 65°C for 1-2 hours to remove genomic
549 DNA, and concentrated using the RNeasy RNA cleanup kit (Qiagen).

550 **16S rRNA gene sequencing and analysis**

551 16S rRNA gene sequence data was generated from 143 samples (**Table S1**). The
552 V3 region of the 16S rRNA gene was amplified using the universal primers (341F: 5'-
553 CCTACGGGRSGCAGCAG-3' and 515R: 5'-TTACCGCGGCKGCTGVCAC-3')
554 (Klindworth *et al.*, 2013). Two-step PCR reactions (modified from Berry *et al.*, 2011)
555 were conducted in 25 µl volume contained 0.5 µM MgCl₂, 0.2 mM deoxynucleotide, 0.2
556 µM each primer and 1U of Phire Hot Start II DNA Polymerase (Finnzymes Thermo
557 Fisher Scientific). The template was amplified using non-barcoded PCR primers for 20
558 cycles, followed by 1:50 dilution of the PCR product and 10 additional cycles of
559 amplifications with barcoded PCR primers. The thermal program consisted of an initial
560 95 °C denaturation step for 4 min, a cycling program of 95 °C for 30 s, 52 °C for 30 s,
561 and 72 °C for 60 s, and a final elongation step at 72 °C for 7 min. Reverse primers were
562 barcoded with specific IonXpress sequence to identify samples. PCR products were
563 purified using QIAquick Gel Extraction Kit (Qiagen), quantified using Quantifluor
564 dsDNA System (Promega), pooled at equimolar concentration and sequenced using an
565 Ion Torrent PGM system on a 316 chip with the ION Sequencing 200 kit as described in
566 Sanschagrín and Yergeau (2014).

567 V3 region 16S rRNA sequences were analyzed using MOTHUR (Schloss *et al.*,
568 2009). Sequences with an average quality of <17, length <100 bp or that did not match
569 the IonXpress barcode and both the PCR forward and reverse primer sequences were
570 discarded. Potential chimeric sequences were identified using UCHIME (Edgar *et al.*,

571 2011) and discarded. Verrucomicrobia sequences were identified by taxonomic analysis
572 in MOTHUR.

573 **Metagenome sequencing, assembly, annotation and binning**

574 DNA sequencing of 24 samples (**Table S1**) was performed at the Department of
575 Energy Joint Genome Institute (JGI) (Walnut Creek, CA, USA) on the HiSeq 2500-1TB
576 (Illumina) platform. Paired-end sequences of 2×150 bp were generated for all libraries.
577 A combined assembly of all raw reads was generated using MEGAHIT
578 (<https://github.com/JGI-Bioinformatics/megahit>) with kmer sizes of 23,43,63,83,103,123.
579 Gene prediction and annotation was performed using the DoE JGI IMG/M functional
580 annotation pipeline (Markowitz *et al.*, 2014). Metagenomic binning based on
581 tetranucleotide frequency and differential coverage was performed with MetaWatt
582 version 3.5.2 (Strous *et al.*, 2012) using the following settings: base pair cutoff of >2000
583 on the co-assembly, and a relative weight of binning coverage of 0.75. The identity of
584 bins was assessed using the phylogenetic analysis of a concatenation of single copy core
585 genes implemented in MetaWatt, MAFFT aligner and FastTreeMP for tree inference
586 (Price *et al.*, 2010). Genome completeness and contamination was assessed with CheckM
587 (Parks *et al.*, 2015), which relies on pplacer (Matsen *et al.*, 2010), prodigal (Hyatt *et al.*,
588 2012) and HMM (Eddy, 2011). Fourteen MAGS with low contamination (<5%) and
589 substantially or near-complete (>70%) were obtained, as described by the bin quality
590 terminology proposed in Parks *et al.* 2015. Despite a completeness value of 58%, we also

591 included Opi-242, given its abundance during the ice-covered period of Lake
592 Simoncouche.

593 **Ecological association of Verrucomicrobia bins**

594 We performed a canonical constrained analysis (CCA) using the *vegan* package
595 (Oksanen *et al.*, 2017) in R (R Core Development Team, 2016). The 20 environmental
596 variables measured were tested for normality using a Shapiro test, and transformed using
597 *powerTransform*, followed by a Box-Cox transformation (*bcPower*), using the *car*
598 package (Fox *et al.*, 2016) for all variables for which $p < 0.05$ in the Shapiro test. A linear
599 model was fitted to each pair of variable, and we removed variables for which the
600 correlation coefficient was greater than 0.70 from the CCA. In brief, we performed the
601 CCA using 14 of the environmental variables, and the complete species matrix containing
602 24 samples and 54 MAG.

603 **Phylogeny of Verrucomicrobia**

604 A concatenated gene phylogeny tree was created based on 4 of the 5 genes as in
605 He *et al.* (2017), in which the 15 Quebec MAGs were put in the contact of the 19 MAGs
606 from Lake Mendota and Trout Bog (Wisconsin), and 7 MAGs from ice-covered Lake
607 Baikal (Cabello-Yeves, Zenskaya, *et al.*, 2017) and 17 MAGs from the Tous and
608 Amadorio reservoirs (Cabello-Yeves, Ghai, *et al.*, 2017). Because none of the 15 MAGs
609 in our study contained the DNA polymerase III beta subunit (but had other subunits such
610 as alpha), the concatenated gene phylogeny contained instead the following four genes:

611 TIGR01391 (DNA primase), TIGR01011 (Small subunit ribosomal protein S2),
612 TIGR00460 (Methionyl-tRNA-formyltransferase), and TIGR00362 (Chromosomal
613 replication initiation factor). 17 out of the 18 MAGs from the Tous and Amadorio
614 Reservoirs had these genes. To create the phylogenetic tree, sequences for each gene
615 were aligned using MUSCLE (Edgar, 2004) and a maximum-likelihood tree was created
616 for each individual gene to ensure that these proteins were conserved and represented the
617 phylogenetic relationships between groups, before concatenation of the sequences using
618 Mesquite (Maddison & Maddison, 2017). The confidence score of each amino acid
619 position in the multiple sequence alignment was calculated using ZORRO (Wu *et al.*,
620 2012), and amino acid position that had a score below 0.5 was manually deleted using
621 Mesquite (Maddison & Maddison, 2017) to obtain a more accurate phylogenetic
622 inference. MEGA6.06 (Tamura *et al.*, 2013) was used to generate a maximum likelihood
623 phylogeny of 65 taxa, using a bootstrap of 100 iterations. The substitution model was the
624 Jones-Taylor-Thornton (JTT) model for amino acids. The rates among site were Gamma
625 distribution, with 4 gamma categories. The maximum-likelihood (ML) heuristic method
626 was Nearest-Neighbour Interchange (NNI), and the initial tree was NJ. Finally, the
627 branch swap filter was Very Strong. Despite that four instead of five genes were used, the
628 phylogenetic structure of the tree (and relationships between taxa) is the same as when
629 using 5 genes, as in He *et al.* (2017). In order to include all 18 MAGs from (Cabello-
630 Yeves, Ghai, *et al.*, 2017), an automated concatenated gene phylogeny was created using

631 PhyloPhlan (Segata *et al.*, 2013) (**Figure S2**). The phylogenetic structure of the
632 PhyloPhlan tree was consistent with the manually curated tree (**Figure 3**).

633 **Comparative genomics and functional annotation**

634 The distribution of protein-encoding gene content between genomes was
635 determined using proteinortho (Lechner *et al.*, 2011). Inference of protein function and
636 metabolic reconstruction was based on the IMG annotations provided by the JGI,
637 including KEGG, Pfam, EC numbers, and Metacyc annotations. Metabolic reconstruction
638 was also facilitated by generated pathway genome databases for each MAG using the
639 pathologic software available through Pathway Tools (Karp *et al.*, 2009). In addition, we
640 annotated carbohydrate-active enzymes using dbCan (Yin *et al.*, 2012) and hydrogenase
641 classes using HydDB (Sondergaard *et al.*, 2016).

642 **Metatranscriptomic analysis of Verrucomicrobia gene expression patterns**

643 cDNA library preparation and sequencing of 24 samples (**Table S1**) was
644 performed at the Department of Energy Joint Genome Institute (JGI) (Walnut Creek, CA,
645 USA) on the HiSeq 2500-1TB (Illumina) platform. Paired-end sequences of 2×150 bp
646 were generated for all libraries. The metatranscriptome dataset comprises of 24 samples
647 (6 Croche, 7 Montjoie and 11 Simoncouche). BBMAP ([https://jgi.doe.gov/data-and-
648 tools/bbtools/bb-tools-user-guide/bbmap-guide/](https://jgi.doe.gov/data-and-tools/bbtools/bb-tools-user-guide/bbmap-guide/)) was used to map the raw quality filtered
649 reads to the CDS in each 15 Verrucomicrobia MAG, using the option *min_id=0.97*
650 instead of the default one of 0.7. The RPKB, recruitments per kilo bases of billion reads

651 per sample, was calculated instead of the RPKM (recruitments per kilo bases of million
652 mapped reads) to control for differences in raw reads between samples.

653

654 **ANI and AAI calculations and fragment recruitment**

655 ANI values were calculated with ANICalculator ([https://ani.jgi-](https://ani.jgi-psf.org/html/anicalculator.php)
656 [psf.org/html/anicalculator.php](https://ani.jgi-psf.org/html/anicalculator.php)) using the default settings, but the alignment fraction was
657 too low (median 0.02-0.03 %) to produce meaningful results. Therefore, we assessed
658 genome similarity using AAI values calculated using CompareM
659 (<https://github.com/dparks1134/CompareM>) and the default settings. Fragment
660 recruitment was performed using BBMAP ([https://jgi.doe.gov/data-and-tools/bbtools/bb-](https://jgi.doe.gov/data-and-tools/bbtools/bb-tools-user-guide/bbmap-guide/)
661 [tools-user-guide/bbmap-guide/](https://jgi.doe.gov/data-and-tools/bbtools/bb-tools-user-guide/bbmap-guide/)) and the following settings: *minid*=94, *maxlen*=500,
662 *idtag*=t. To create the heatmap of relative coverage values, we divided the number of
663 reads mapped by the size of the metagenome for each MAG.

664

665 **Data accessibility**

666 The 24 metagenomes and the associated co-assembly can be downloaded at the
667 JGI (IMG Genome ID: 3300010885). The 24 metatranscriptome and the associated co-
668 assembly can be downloaded at IMG Genome ID: 3300013295. MAG data is available at
669 NCBI under the accession numbers XXX-XXX

670

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683

684 **Conflict of Interest Statement**

685 The authors have no conflict of interest to declare.

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- 919
920

921 **Figure titles**

922

923 **Fig. 1. A.** Locations of Lake Croche, Lake Montjoie and Lake Simoncouche in temperate
924 and boreal regions of Quebec, Canada. **B.** Environmental variability of lake samples
925 based on principal component analysis of 21 physiochemical variables. **C.** Relative
926 abundance of Verrucomicrobia in the 16S rRNA amplicon datasets organized by lake and
927 season. I-C and I-F refer to ice-covered and ice-free periods, respectively. Mid-lines
928 represent median, upper and lower boundaries are 25% quartiles, and whiskers represents
929 scores outside the middle 50%.

930

931 **Fig. 2. A.** Concatenated protein phylogeny of sampled lake MAGs and Verrucomicrobia
932 reference genomes. **B. and C.** CCA of the 54 MAGs constrained to 14 environmental
933 variables (arrows). The 24 samples are shown in the ordination space in panel b, whereas
934 the 54 MAGs are shown in panel c. Abbreviations for environmental variables are
935 described in Table S2.

936

937 **Fig. 3 . A.** Concatenated maximum likelihood phylogeny, using 100 bootstraps, of
938 selected Verrucomicrobia MAGs and reference genomes. **B.** Heatmap of the relative
939 MAG abundance based on coverage (blue) and relative gene expression based on the
940 number of transcripts recruited (red) to each MAG. Rectangular boxes represent ice-

941 covered samples, square boxes represent epilimnetic (top squares) and the metalimnetic
942 (bottom squares) samples.

943

944 **Fig. 4. A.** Venn diagrams of ortholog distributions between MAGs in V1 (left panel) and
945 V4 (right panel). **B.** Enzymatic reaction for limonene (top) and cymene (bottom)
946 degradation identified in V1-690. **C.** Gene clusters involved in sulfur oxidation in V1-690
947 and hydrogen oxidation in Opi-474. **D.** The α -oxidation pathway for phytol degradation
948 in Opi-474 and Opi-242. For panels **b** and **e**, E.C. numbers are shown, and shaded boxes
949 represent RPKB during winter. Numbers inside the boxes represent
950 “ScaffoldNumber_Loci”.

951

952 **Fig. 5 A.** Expression of Chth-196 genes involved in flagellar motility. **B.** Expression of
953 Chth-196 genes involved in chemotaxis **C.** Expression of genes in the urease gene cluster
954 in Chth-196 and Pedo-303 **D.** Genes involved in ammonia utilization and degradation in
955 Chth-196 and Pedo-303. Since Chth-196 is present in Montjoie and Simoncouche, but
956 Pedos-303 is only present in Montjoie, only those lakes are shown. Genes for which
957 expression was positive (Reads per kilo bases per billions reads – RPKB) are highlighted
958 with colour.

959 **Tables**

960 **Table 1.** Genome characteristics of the 15 Verrucomicrobia MAGs, recovered from
 961 Quebec lakes metagenomes co-assembly.

Subdiv.	MAG	Sequence length (Mbp) ^a	N50 ^a	Coverage ^b	GC Content (%) ^a	Gene Count ^a	Completeness (%) ^a	Contamination (%) ^a	Estimated bin size (Mbp) ^c
1	V1-33	4.08	49467	11.9	63.31	3225	78.24	0.68	5.18
1	V1-115	4.14	40207	9.9	61.77	3420	82.07	4.55	4.81
1	V1-159	2.63	31999	16.3	62.55	2373	95.92	0	2.74
1	V1-690	2.24	17084	12.3	60.46	2015	70.61	2.86	3.08
1	V1-1361	2.25	11724	23.8	60.29	2055	72.93	0.68	3.06
2	Chth-244	3.24	18974	13	65.1	3122	83.28	2.2	3.80
2	Chth-196	2.79	35078	15.1	58.27	2263	70.16	0	3.98
2	Xiphi-554	1.66	24984	10.4	42.77	1576	95.95	0.71	1.72
2	Xiphi-315	1.42	22457	38.9	54.51	1400	87.73	2.03	1.59
3	Pedos-303	4.47	17418	9	67.77	3693	89.7	2.7	4.85
3	Pedos-1123	2.73	11423	12.6	52.1	2471	95.05	2.59	2.80
3	Pedos-510	2.51	16489	25.8	54.52	2380	82.43	0.34	3.03
4	Opi-242	1.19	34025	63.8	54.32	1116	57.77	0	2.06
4	Opi-128	1.85	18982	107.1	42.47	1684	77.74	0.68	2.36
4	Opi-474	2.81	11480	51.6	65.6	2690	92.17	4.11	2.92

962

963 ^a Calculated using CheckM

964 ^b Calculated using MetaWatt

965 ^c Calculated as (sequence length/completeness)*(100-contamination)

966 **Supplementary Figures and Tables Titles**

967

968 **Figure S1.** Scatter plots of chlorophyll *a*, total phosphorus (TP) and total nitrogen (TN)

969 coloured by strata over the period of the study in the three Quebec lakes.

970

971 **Figure S2.** Concatenated gene phylogeny using PhyloPhlan showing the 15 Quebec

972 MAGs with all currently existing Verrucomicrobia MAGs from Lake Baikal, Trout Bog,

973 Lake Mendota, Tous and Amadorio reservoirs.

974

975 **Table S1.** Description of the 16S rRNA gene, metagenome and metatranscriptome

976 datasets employed in this study.

977

978 **Table S2.** Environmental variables measured in the field and in the laboratory.

979 Abbreviations are the same as those used in Figures 1B and 2B.

980

981 **Table S3.** Pair-wise amino acid identities (AAI) comparisons between all available

982 freshwater Verrucomicrobia MAGs, divided into subdivisions.

983

984 **Table S4.** Fragment recruitment results of 15 MAGs against 22 freshwater metagenomes.

985

986 **Table S5.** Expression values (RPKB) for GHs in each of the 15 MAGs.

Figure 1

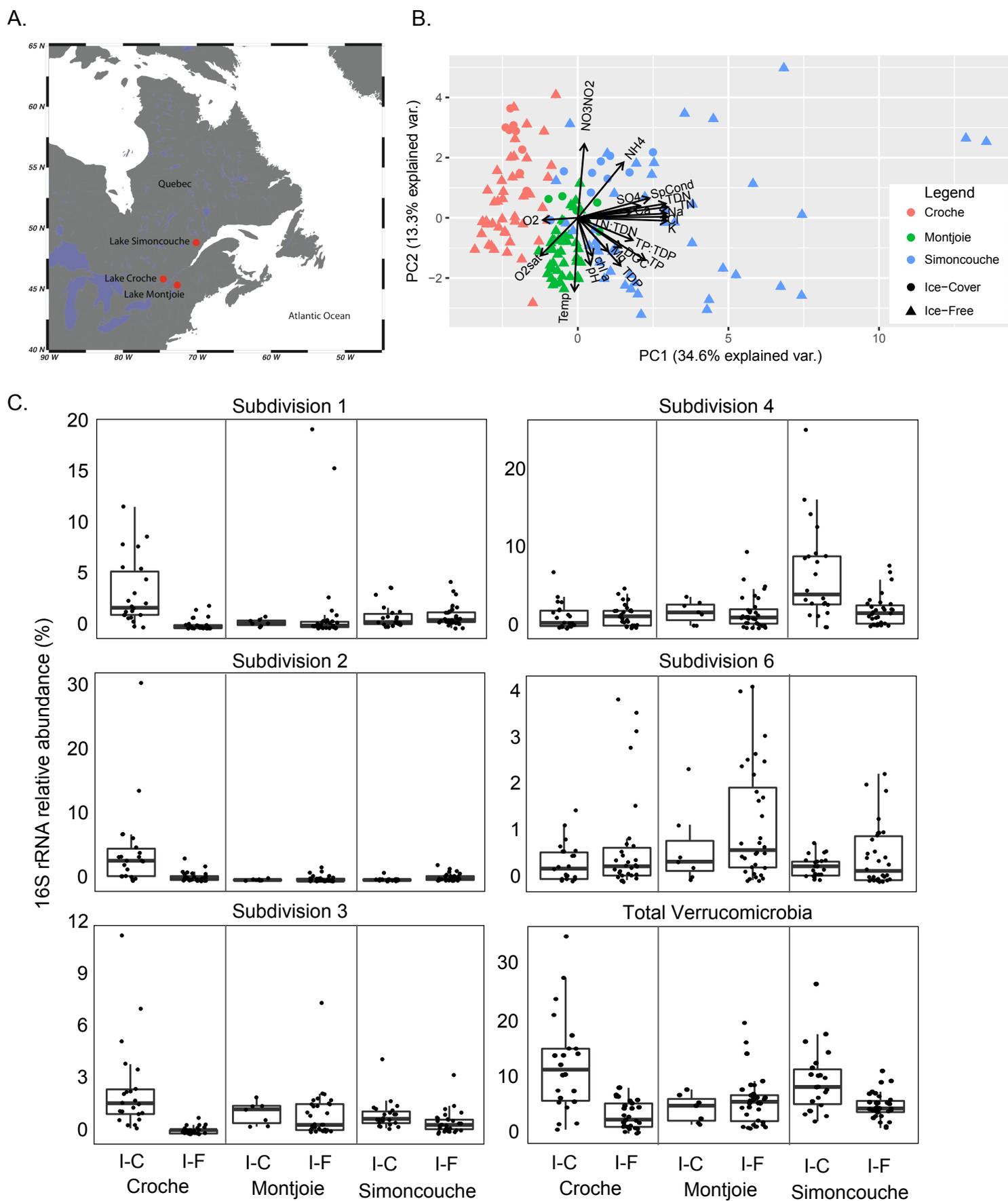
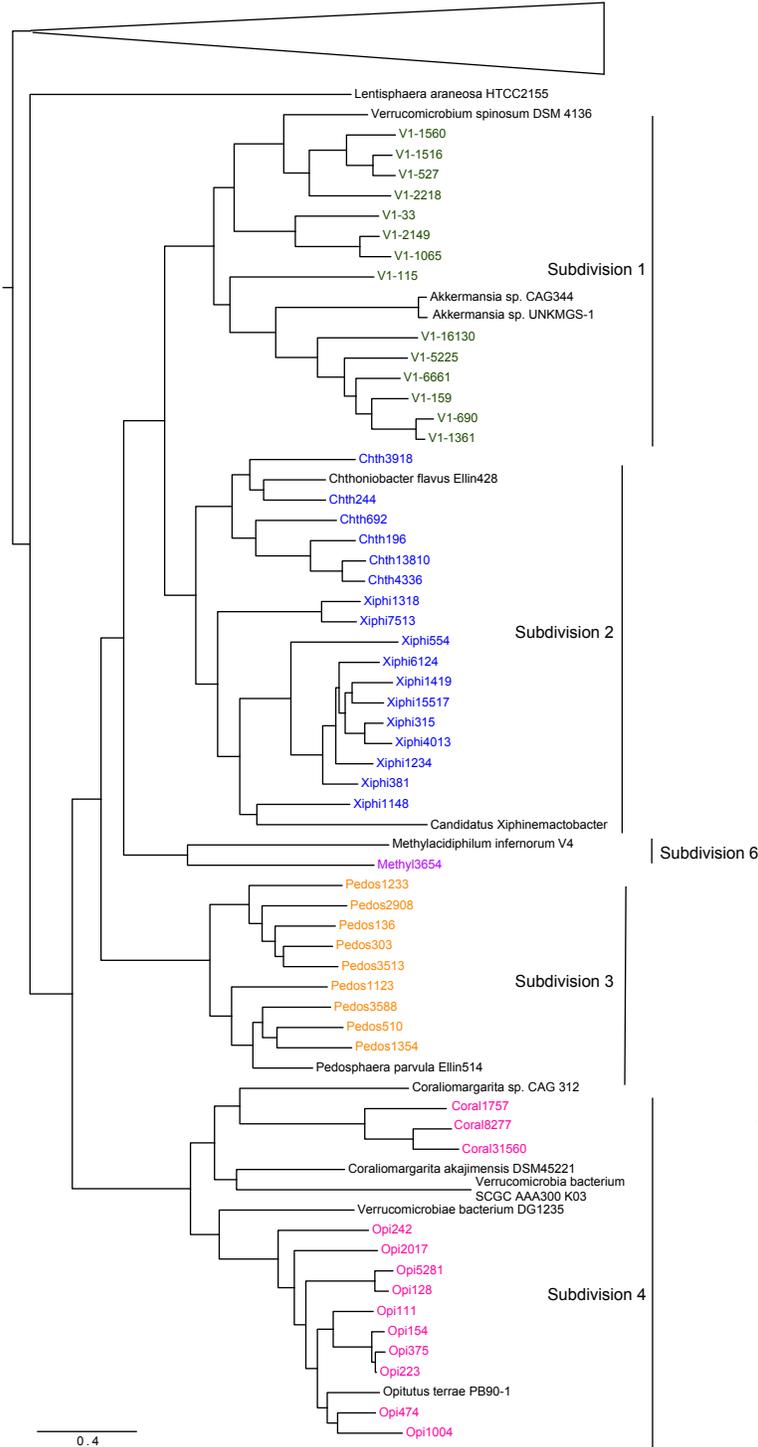
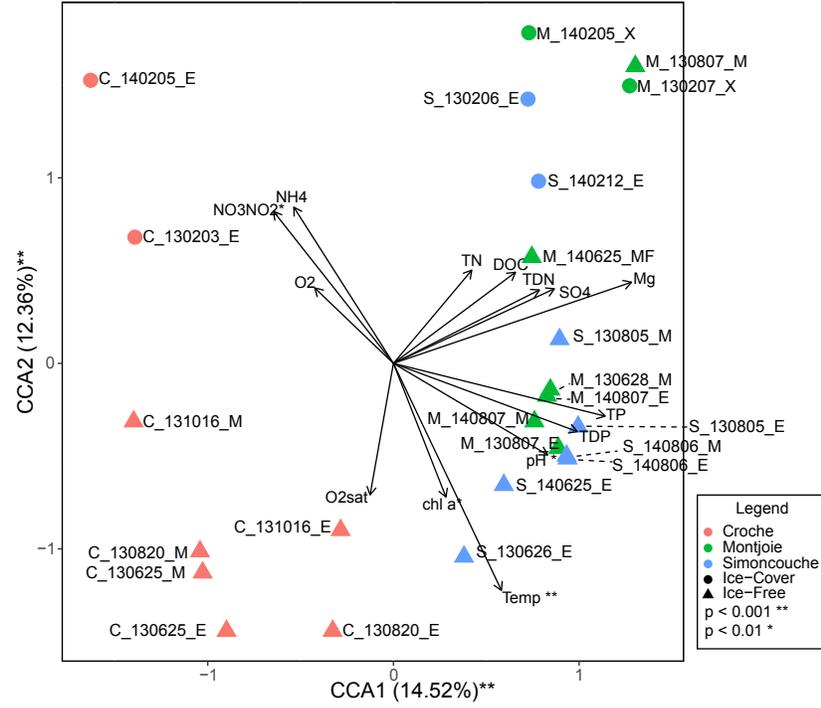


Figure 2

A. Phylogeny of Verrucomicrobia MAGs



B. CCA: Sites and environmental variables



C. CCA: MAGs ecological associations

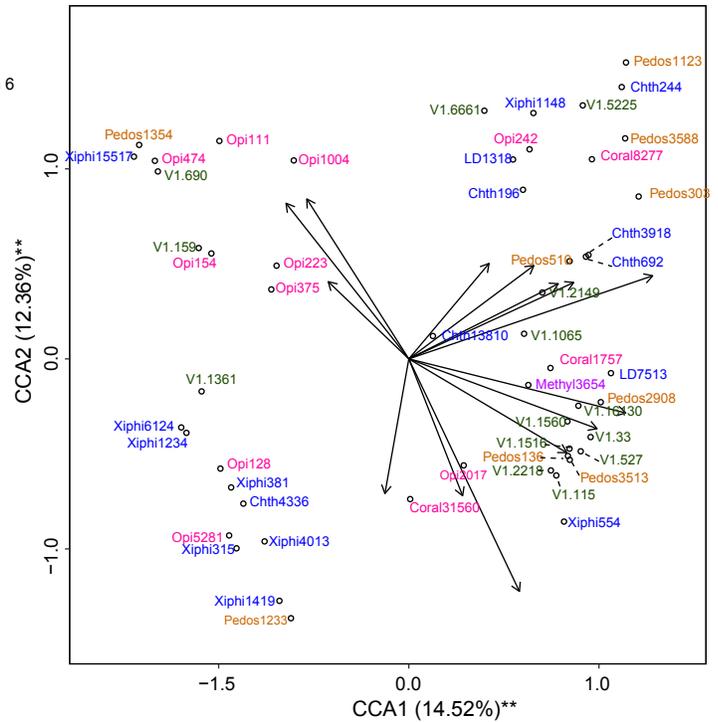
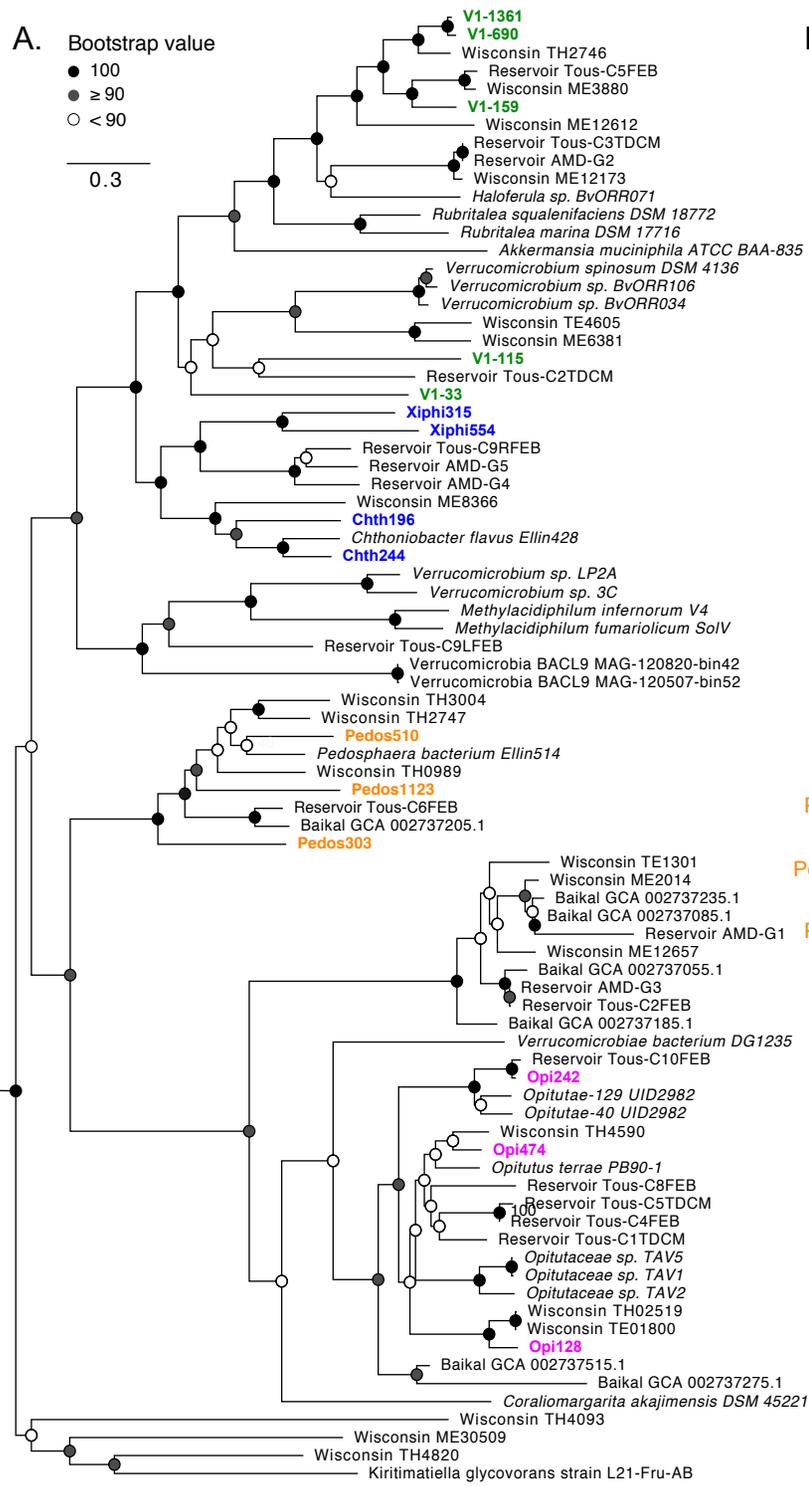


Figure 3



B.

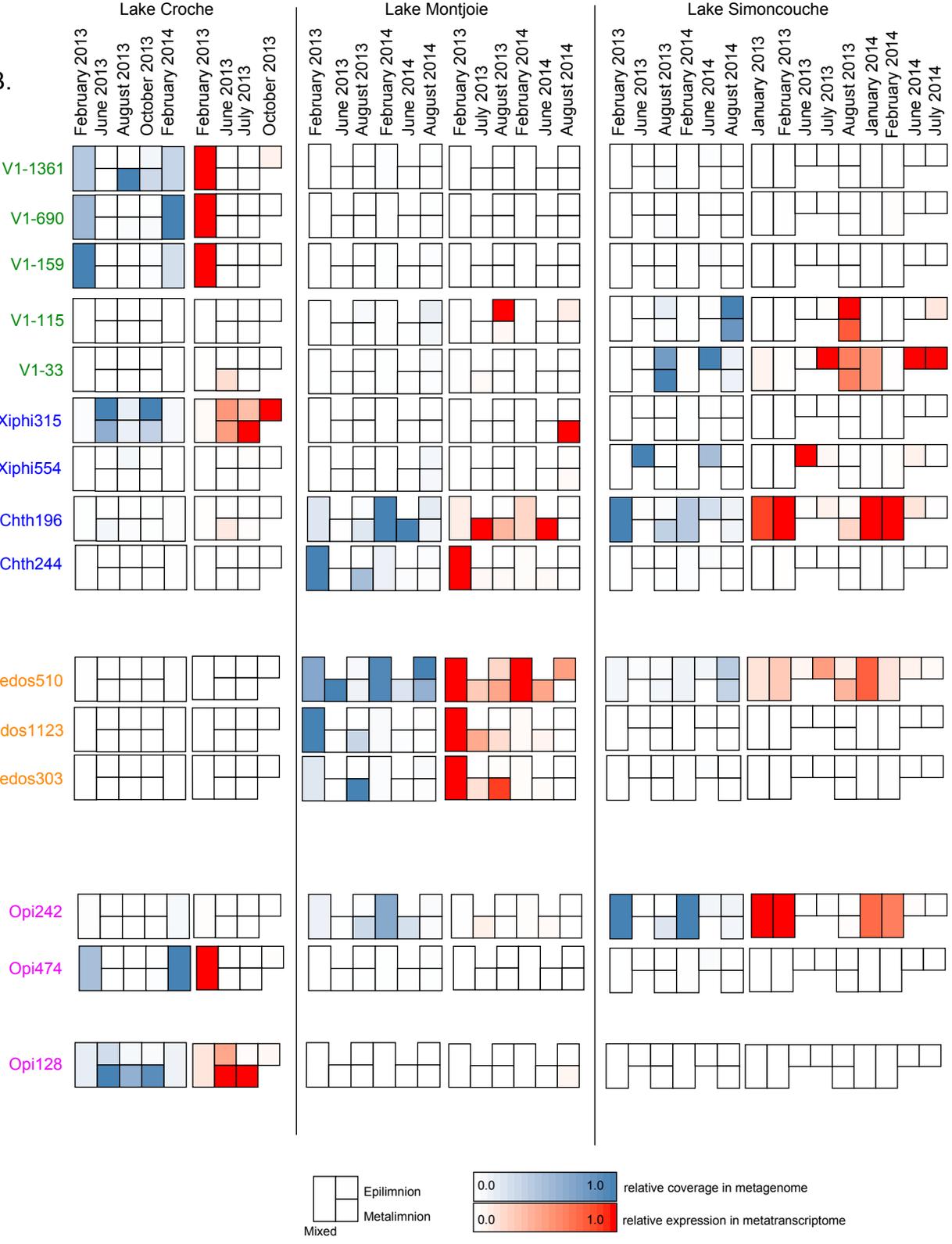
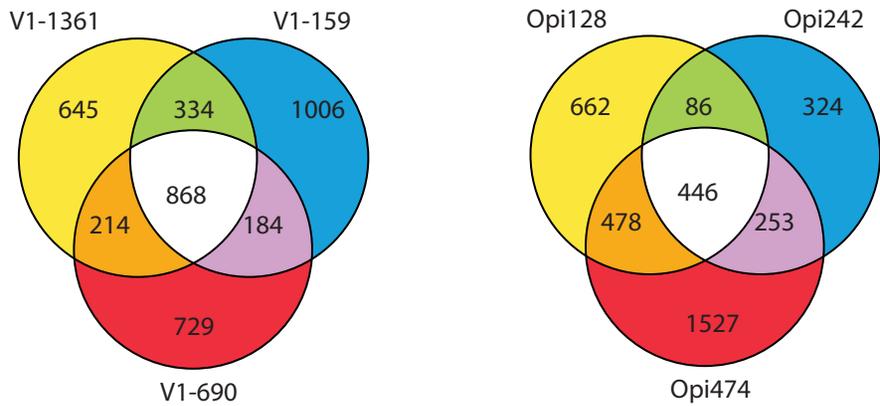
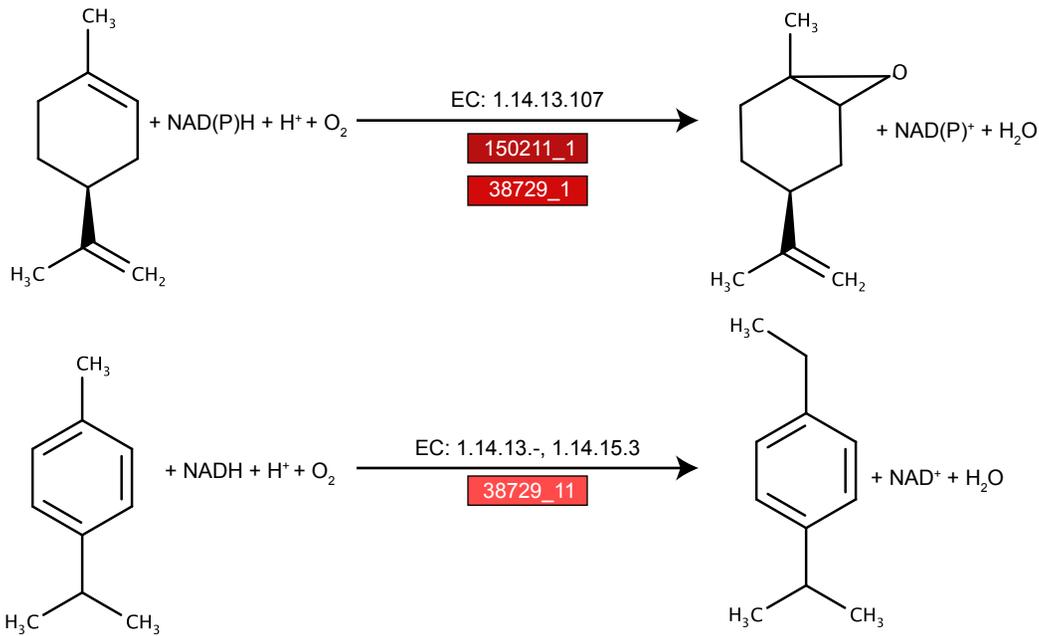


Figure 4

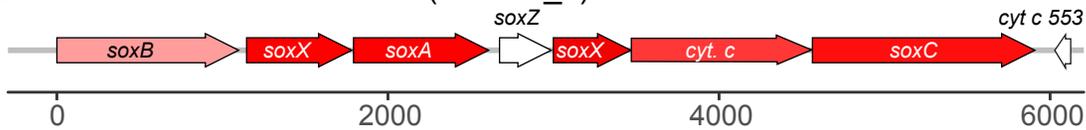
a.



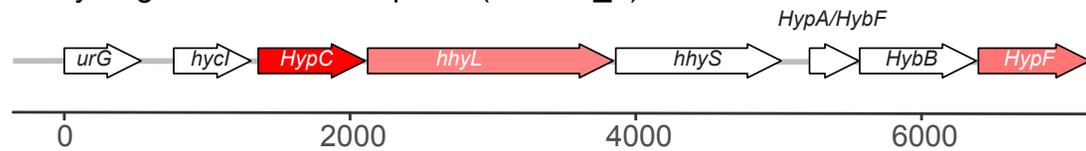
b. V1-690



c. Sulfur oxidation in V1-690 (138384_3)



Hydrogen oxidation in Opi474 (115945_4)



d.

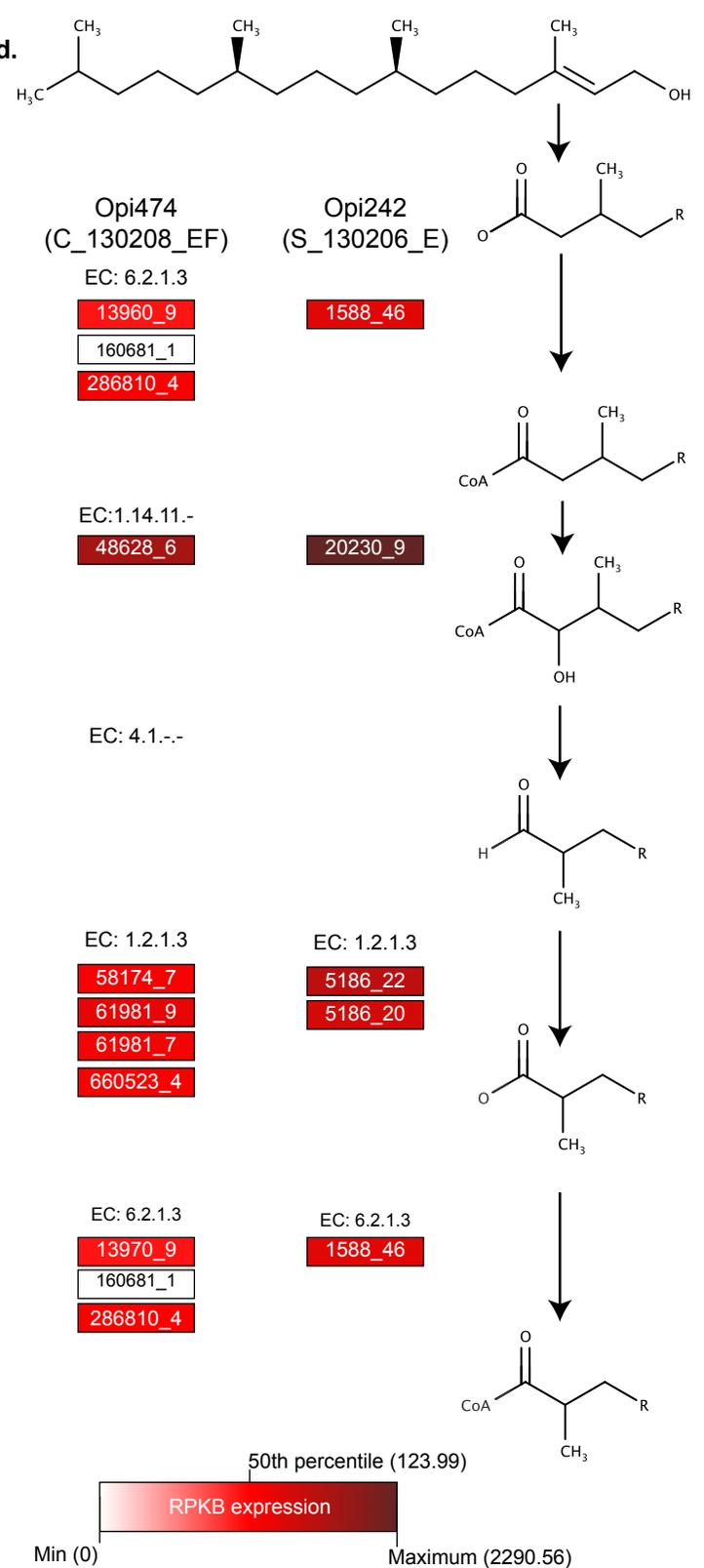
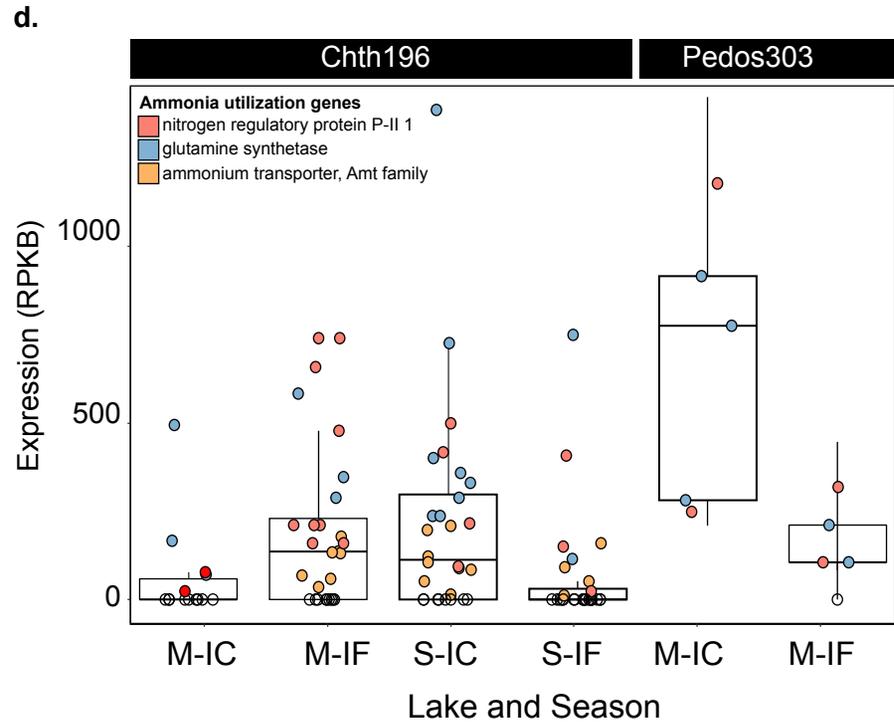
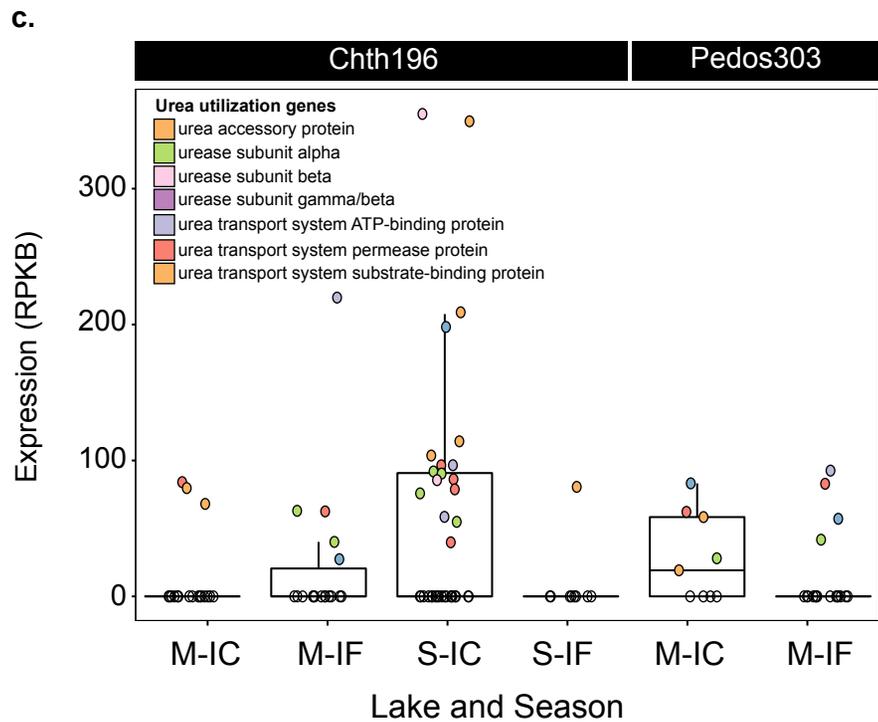
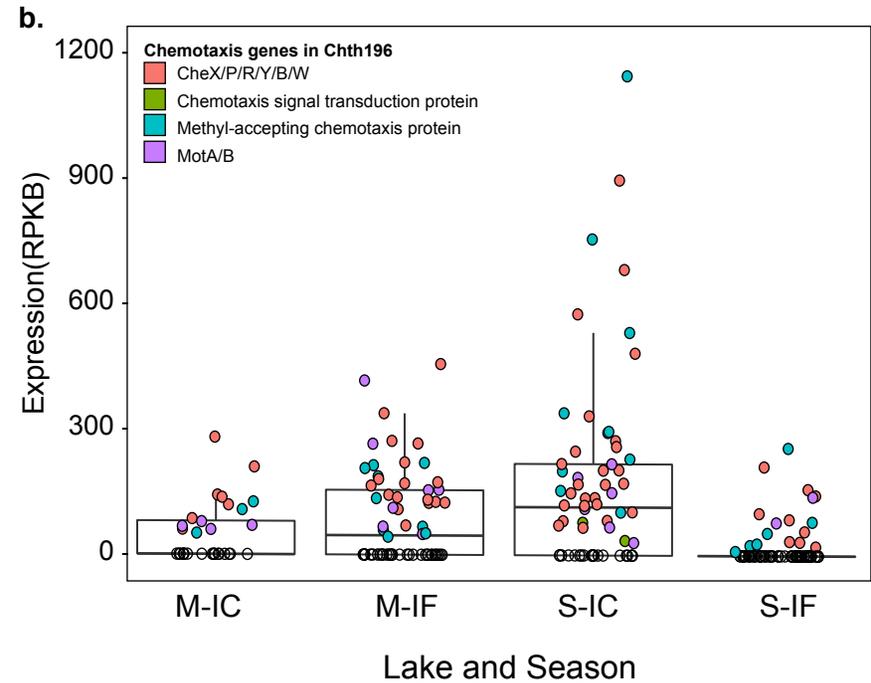
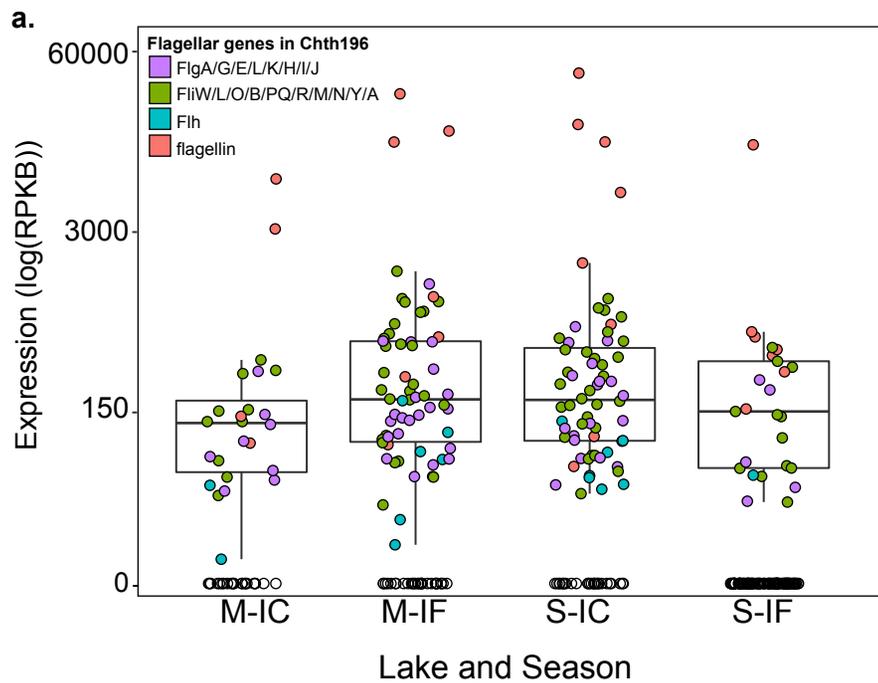


Figure 5



Supplementary Figure 1

